

1. INTRODUCTION

 Bees are considered to be some of the most important pollinators of both wild and cultivated plants (Michener 2000), responsible for pollinating at least 60% of the nearly 1500 species cultivated by humans (Garófalo 2009). Stingless bees have attracted the attention of researchers interested in the effects of deforestation and forest fragmentation on pollinators (Liow et al. 2001, Cairns et al. 2005, Villanueva-Gutiérrez et al. 2005, Ricketts et al. 2008, Brosi 2009, Brosi et al. 2007, 2008, Freitas et al. 2009, Fierro et al. 2012). The main conclusion from this work is that tropical deforestation affects greatly the abundance, diversity and composition of stingless bees, and that deforestation could have serious consequences for the pollination and reproduction of both native and cultivated plants. Few studies exist, however, from the Brazilian Amazon, in spite of its putative high diversity of bees (Oliveira et al. 1995, Oliveira 2001, Dick 2001, Brown and Albrecht 2001).

 In Brazil, stingless bees comprise one of the country's most species-rich groups, with 192 recorded species (Silveira et al. 2002); the actual number of species is likely much higher, considering how poorly sampled bees are in the Brazilian Amazon (Overal 2001, Baccaro et al. 2008). There is a nearly forty-year gap in the generation of knowledge of these bees in the region. That gap runs from the last study of Adolpho Ducke (Ducke 1925), who pioneered the study of these bees in the region, to expeditions that began in the region in 1963 (Camargo 1994). There are also major spatial gaps in our knowledge. Most of the work cited above was carried out along the margins of major rivers (Camargo 1994) and near major urban areas (Oliveira et al. 1995, Oliveira 2001), with almost no work in the immense areas between rivers, near headwaters or in

 mountainous areas (Oliveira et al. 2010). While our knowledge of these bees is relatively sparse in the region, deforestation due to development of infrastructure, mining, and agricultural colonization continues throughout the Amazon, in spite of notable decreases in Brazil and several other Amazonian countries from 2005-2010 (Colombia, French Guiana, and Peru actually saw increases in deforestation rates) (PRODES 2011, RAISG Amazonian Network of Georeferenced Socio-Environmental Information 2013).

 The landscape of the state of Rondônia, Brazil, the focus of the present study, is emblematic of the effects of development in the Brazilian Amazon on forest cover and biodiversity. Up until the 1960s, the forests of Rondônia were relatively intact, the main forest type being "tropical moist forest" in the northern half of the state and "subtropical moist forest" according the Holdridge Life-Zones map (International Institute for Applied Systems Analyses (IIASA 1989). Agricultural colonization projects were established there beginning in the 1960s. Ouro Preto do Oeste (hereafter "Ouro Preto") was the first colonization project established in Rondônia, it is one of thirteen meso-regions of the state surveyed for the present study, and unlike the other regions, it was chosen for intensive monthly sampling during the study period for the following reasons. Ouro Preto's landscape is a microcosm of the range of deforestation landscapes found across the state, with environments representing some of the longest settlement history in the state and some much more recent, leading to areas ranging from very high to very low levels of deforestation, respectively (Figure 1). In general, then, farm lots within 16 km of the main BR-364 highway bisecting the state from southeast to northwest had very little forest remaining by 1996, when the present study was conducted, in contrast, there is much more forest remaining on farm lots in Mirante da Serra near the indigenous reserve

 Statistical analyses involved a number of different techniques to explore the relationships among deforestation, species richness and composition at the above spatial scales. These included scatterplots, ordination (conducted using Canoco 5), Ordinary Least Squares regression, and by comparing the summary statistics of all locations within particular ranges of deforestation level using boxplots. Regarding ordination, detrended correspondence analysis of species composition across sample locations indicated the use of linear methods, so redundancy analysis was used for all ordination. Before conducting the ordination, rare species were removed from the dataset by excluding species that appeared in less than 5% of sample sites. For analyses involving just the Ouro Preto meso-region, this left 62 species, and for the state-wide dataset, this left 63 species.

2.2. Choice of collection locations

 Fieldwork was sponsored by the Second Approximation Project of the Socio- Economic-Ecological Zoning of Rondônia, funded by the World Bank and executed by Tecnosolo/DHV Consultants. Selection of locations for sampling species composition and richness had to take into account the main access ways to the priority areas of the research for the zoning exercise: the federal highways BR-364, and 429, secondary roads, and the Ouro Preto and Cautário Rivers (Figure 1). Collections were most often done where access was easiest. Factors such as heavy rains, poor road conditions, and the need to obtain permission from landowners to enter properties often limited access. Examination of the number and spatial distribution of sample locations throughout the state, however, suggests that the data are unprecedented in spatial coverage in comparison to other stingless bee surveys, and they are representative of the state and its

3. RESULTS

3.1. Overall species composition in Rondônia

3.4.1. Species composition

 Reserva Extrativista do Rio Ouro Preto, the Estação Ecológica Cuniã, and the Parque Estadual Guajará-Mirim.

3.5. Micro-regional analysis

3.5.1. Species composition

Sixty-one species (62%) were found at least once, no matter the level of

deforestation, suggesting these bees may have some level of tolerance to deforestation

and fragmentation of the landscape (supplementary materials, Table C). In contrast,

twenty-seven species (27%) were the only species absent from highly deforested areas

(80-100% deforestation) suggesting a susceptibility to deforestation. Redundancy

analysis showed statistically significant explanation of the variation in species

composition using dummy variables for the meso-region of each sample site and the .5

km deforestation variable as environmental variables (adjusted explained variation 3.5%;

Monte Carlo permutation test results (499 permutations): first axis (pseudo F-ratio=4.4,

297 p=.004) all axes (pseudo F-ratio=1.5, p=.002)(Figure 6). The .5 km deforestation variable

by itself was insignificant (results not shown).

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- 300 <Fig 6 about here>
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3.5.2. Species richness

 A scatterplot of the percent of area deforested within 0.5 km of the location of each sample site and species richness for all locations across Rondônia (Figure 7a) shows a similar situation to that found for the Ouro Preto meso-region (Figure 3). Results for the

 Rondônia can be said now to have 110 species, known as one of the most species-rich and sampled in the tropics with respect to stingless bees. A discussion of potential taxonomic issues with our list of species can be found in supplementary materials. The large number of substrates examined for collecting in this study likely helped maximize the number of species found in each location, with nearly 80% of all species captured from flowers, honey baits, and on skin (collecting sweat) (Table A, supplementary materials).

 The evidence for the impact of colonization and subsequent deforestation on stingless bees was most visible from the more intensive yearlong collection in the Ouro Preto meso-region. More species were found there than in any other meso-region, likely due to the large number of samples taken throughout the year. On the left side of figure 3a are data from collection sites that are closest to the BR-364 and the urban center of Ouro Preto. Thus, they are within the oldest areas of colonization, ones that consequently are the most deforested as well. From km 4 to km 32, the number of species tends to rise, indicating that with distance from the BR-364, the species richness rises. From km 32 to 60, however, richness decreases somewhat, which could be attributed to deforestation and urban impact, because km 40 is the center of the urban area of Nova União, and km 60 is Mirante da Serra. Past km 60 to km 84, the end of colonization and the border of the indigenous reserve and national park, richness tends to rise again, with three sites of very high richness at km 84. There was, however, a statistically significant relationship between species richness and deforestation within .5 km of the sample locations (Figure 3 a and b). The redundancy analysis bi-plot in figure 3d shows the 15 best-fitting species, with all but *Melipona fuliginosa* strongly negatively correlated with the deforestation

 however, was not related to any abnormal weather. One possible explanation for low richness is that in very large preserved areas, the sampling methods used simply were inadequate to detect the existing diversity. In smaller forested patches, the chances may be higher to find greater diversity. M.L. Oliveira (personal observation) found a similar situation when sampling orchid bees (Euglossini) in the region of Manaus, and J. M. F. Camargo (Unpubl. data) remarked that stingless bees in the Amazon express very patchy distribution, with many species concentrated in few places, leaving some larger areas with low diversity within forested areas. The redundancy analysis at the meso-region level revealed some important results. The .5 km deforestation variable alone was insignificant in explaining species composition, unlike the case when only the Ouro Preto meso-region alone was analyzed. An ordination that included dummy variables for the meso-region of each sample site, however, was statistically significant (Figure 6), and showing the 20 best-fitting species

in the species-environmental variables bi-plot showed a group of species highly

negatively corrected to deforestation, indicating possible susceptibility to deforestation.

Finally, the local-scale analysis showed no statistically significant difference in species

richness, but redundancy analysis of species composition showed a statistically

significant difference between closed and open canopy environments. An examination of

the 20 best-fitting species along the horizontal axis allowed for identification of

potentially susceptible species (Figure 8).

 To sum up the effects of agricultural colonization and subsequent deforestation, we look to the redundancy analyses to identify particular species and groups that appear most strongly affected (Figures 3, 6, and 8). We start by selecting the best-fit species in each of

 the three ordinations, and from those selecting the most negatively correlated with the deforestation variable. We then list those species as a first cut of the most affected by deforestation. Relative susceptibility within this list can be further determined by seeing which species appear most frequently in the list across the ordinations.

 Table III marks with an "x" the species that appear in the corresponding ordinations according to the above rules. It bears repeating that these ordinations are based on datasets that exclude rare species, so they are all species that are widespread in Rondônia and independent of the region in which they were found showed negative correlation with the .5 km deforestation variable and were mainly found in closed canopy environments. Clearly, not all species known to prefer cavities in live trees are in our list of species most likely affected by deforestation. This may be because they were too rare in our survey to be included in the redundancy analyses or their presence has yet to be affected by deforestation. For any species to persist in an area undergoing deforestation, 433 stingless bees must survive the physical destruction that occurs during tree felling, the initial burning of the trees, and subsequent burns, sometimes annual, especially in areas of cattle pasture. Bees that persist must have the ability immediately to rectify and rebuild nest architecture and then survive repeated burning and predation. As an example *M. seminigra abunensis* and *M. grandis* appears able to do this better than other *Melipona* species in the Ouro Preto meso-region, or perhaps these larger *Melipona* species respond to disturbance over a longer period of time (Brown and Albrecht 2001). We do not know whether the species and colonies found in open areas moved into those areas after disturbance, or whether they survived the disturbance. In the long term, species must survive potential isolation and inbreeding.

 There appears to be a consensus that stingless bees are essentially a forest group, but as generalists, they are able to forage away from their nests into disturbed environments in many cases, as supported by our analysis of species composition in open and closed canopy environment. There are likely to be several species that can survive quite well in disturbed environments, the classic species in this case being *Tetragonisca angustula*; it is very common in disturbed areas in the Americas (Oliveira 2001, Fierro et al. 2012), and is even well adapted to urban environments where it is commonly found in buildings, wooden posts, walls, and it is widely managed for honey production. On the whole, it seems prudent to follow Brosi et al. (2008) and Brosi (2009) who recommend preservation of forest fragments wherever possible to maximize the possibility of colony survival.

 Numerous questions remain for future research. It seems clear that stingless bee sampling is currently ineffective in very large, forested areas (Oliveira 2001). The bees may simply be easier to find and capture in deforested areas, so it would be helpful to understand better the spatial pattern of foraging by bees, perhaps by experimentation with managed colonies in forested areas and accompanying studies of pollen types found in honey throughout the year to determine the relative contributions of environments of various disturbance levels to colony survival. We also have little idea of how colonies survive the process of deforestation and subsequent burning of agricultural plots by settlers. It would seem plausible that maintenance of forest fragments is essential for stingless bee conservation, but future studies should attempt to determine what the minimum size and ideal spatial configurations are for species conservation.

 It is our hope that this work brings greater attention to this group of bees as a resource that provides pollination services for both native and non-native plants and crops (Santos and Absy 2010, Rech and Absy 2011a and 2011b). These bees are affected by agricultural settlement and deforestation and we have an opportunity to plan for their conservation as areas undergo development in years to come.

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FIGURE CAPTIONS

 Figure 1. Map of study locations and the meso-regions studied in the state of Rondônia, Brazil.

Figure 2. Stingless bee species accumulation curve, across 73 collection locations, during

one year of sampling in Ouro Preto do Oeste meso-region, Rondônia, Brazil.

- Figure 3. (a) Scatterplot of stingless bee species richness in each sample location vs. the
- distance of the location from the BR-364 highway, Ouro Preto meso-region, Rondônia,
- Brazil. (b) Number of species found in each location vs. the percent area deforested
- within 0.5 km of sample locations. Both linear regression lines are significant (p≤0.0001).
- (c) Bi-plot of species and .5 km deforestation variable, all species. (d) Bi-plot showing

only the 15 best-fitting species (DolcLong-*Dolochotrigona longitarsis*, FrieSilv-

Frieseomelitta silvestrii, MelBr-*Melipona brachychaeta*, MelpFulg – *Melipona*

fuliginosa, MelpSchw-*Melipona schwarzi*, PartVici-*Partamona vicina*, PlbAffMn-

Plebeia aff. *minima*, PlebKerr-*Plebeia kerri*, PlebMarg-*Plebeia margaritae*, PlebeSp1-

Plebeia sp. 1, PlebeSp2-*Plebeia* sp. 2, SchwTimd-*Schwarzula timida*, TetrgSpN-

- *Tetragona* sp.n., TrigGrae-*Trigonisca graeffei*, TrigHirt-*Trigonisca hirticornis*,
-
- 627 Figure 4. Comparison of stingless bee richness during wet $(w; n=38)$ and dry $(d; n=35)$
- seasons in Ouro Preto do Oeste meso-region in Rondônia state, Brazil. Shaded areas of

box plots are 95% confidence intervals around the median.

Figure 5. Comparison of stingless bee species richness per sample location across type of

meso-region in Rondônia state, Brazil. Shaded areas are 95% confidence intervals around

the median. (AP n=54; newer n=31; older n=102). AP=Area of preservation;

newer=meso-regions receiving greatest influx of migrants 1981 and later; older=meso-

regions receiving greatest influx of migrants 1980 and earlier.

Figure 6. (a) Redundancy analysis species-environmental variables bi-plot of samples

from all meso-regions (63 most common species, .5 km deforestation variable, and

dummy variables for all meso-regions). (b) Plot from same analysis as a, but showing

only the 20 species with the best fit; species arrows with dotted lines are the most

negatively correlated to the .5 km deforestation variable. DolcLong-*Dolichotrigona*

longitarsis, LeurMuel-*Leurotrigona muelleri*, MelBr-*Melipona brachychaeta*,

MelpSchw-*Melipona schwarzi*, NannMeln-*Nannotrigona melanocera*, PartTest-

Partamona testacea, PlbAffMn-*Plebeia* aff. *minima*, PlebeSp1-*Plebeia* sp. 1, PlebeSp2-

Plebeia sp. 2, PtlLr-*Ptilotrigona lurida,* ScaptSp2-*Scaptotrigona* sp. 2, TetrAngs-

Tetragonisca angustula, TetrDors-*Tetragona dorsalis*, TrigAmaz-*Trigona amazonensis*,

TrgCilCl-*Trigona cilipes*, TrigChan-*Trigona chanchamayonensis*, TrigCras-*Trigona*

crassipes, TrignSp2-*Trigona* sp. 2, TrignSp3-*Trigona* sp. 3, TrigTruc-*Trigona truculenta*,

Figure 7. Relationship of stingless bee species richness and deforestation across the state

of Rondônia, Brazil. (a) Scatterplot of percent of area deforested within 0.5 km of

sample points and species richness. Slope of ordinary least squares regression line is

significant at p≤0.0001. (b) Summary statistics and box plots of distribution of data for

- 654 groups of deforestation level (low $-0<10\%$; medium $-10<80\%$; high $-80-100\%$).
- Shaded areas of box plots are 95% confidence intervals around the median.
-
- Figure 8. Redundancy analysis species-environmental variable bi-plot of 401 sample sites
- of open vs. closed canopy collections. (a) all species, (b) 20 best-fitting species on
- horizontal axis. DolcLong-*Dolichotrigona longitarsis*, FrieTric-*Frieseomelitta*
- *trichocerata*, MelpGran*-Melipona grandis*, ParatSpN-*Paratrigona* sp. n. aff. *lineata*,
- PartNham-*Partamona nhambiquara*, PartVici-*Partamona vicina*, PlbAffMn-*Plebeia* aff.
- *minima*, PlebeSp1-*Plebeia* sp. 1, PlebeSp2-*Plebeia* sp. 2, PlebVari-*Plebeia variicolor*,
- ScauLati-*Scaura latitarsis*, ScauLong-*Scaura longula*, TetrClav-*Tetragona clavipes*,
- TetrGoet-*Tetragona goettei*, TetrAngs-*Tetragonisca angustula*, TetrWeyr-*Tetragonisca*
- *weyrauchi*, TrigAmaz-*Trigona amazonensis*, TrigCras- *Trigona crassipes*, TrigChan-
- *Trigona chanchamayoensis*, TrignSpN*-Trigona* sp. n*.*
-
- Figure 9. Box plot and summary statistics of stingless bee species richness with sub-
- locations grouped by landcover type in Rondônia state, Brazil. Shaded areas of box plots
- are 95% confidence intervals around the median.
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Place	Number of species	area (km ²)	reference				
Madagascar		587,041	Camargo and Pedro (1992)				
New Guinea		462,840	Camargo and Pedro (1992)				
Australia	$8 - 10$	7,692,024	Camargo and Pedro (1992)				
Central Sumatra	24	473,000	Salmah et al. (1990)				
Africa	50	30,221,532	Camargo and Pedro (1992)				
Brazil: Manaus	54	11,401	Oliveira et al. (1995)				
Brazil: Roraima	56	224,299	Oliveira et al. (2010)				
French Guiana	69	83,846	Roubik (1989)				
Brasil:Rondônia	93	237,576	This paper				
Brasil:Ouro	82	3,150	This paper				
Preto do Oeste							

673 Table I. Comparison between the richness of stingless bees recorded in Rondônia state 674 (Brazil) and other places in the Tropics.

677 678 Table II. Summary statistics for stingless bee species richness (r) within each meso-

679 region in Rondônia state, Brazil.

ັ Meso-region	Sample	Total r	Mean r	Median r	StdDev r	Min r	Max r
	locations						
Campo Novo		43	18.2	16	7.40	11	27
Costa Marques		49	19.4	19	4.77	14	26
Extrema	5	52	21	20	7.87	10	30
Machadinho-Cujubim	10	61	21.1	20	10.86	5	37
São Miguel do Guaporé	6	54	18.8	17	10.26	5	36
Estação Ecológica Cuniã	14	35	8.9	8.5	3.75	$\overline{\mathcal{E}}$	15
Parque Estadual Guajará-Mirim	18	63	12.4	10.5	5.75	$\overline{4}$	26
Reserva Extrativista do Rio Cautário		41	16.3	19	5.41		21
Reserva Extrativista do Rio Ouro Preto	15	56	12.9	13	7.96	$\overline{2}$	30
Chupinguaia	8	37	12.1	12	7.62	4	22
Linha D	6	47	14.5	17	8.60		25
Ouro Preto	73	82	13.9	13	7.77		35
Pimenteiras	4	31	11.3	11.5	8.22	2	20
Porto Velho BR-364	\mathcal{E}	22	9.7	9	9.02		19
Rolim de Moura	8	47	13.3	15	5.15	6	21

680 Light shading=newer settlement; dark shading=area of preservation; no shading=older settlements.

681

- 683 Table III. List of species appearing most negatively correlated to deforestation. "X"
-

684 marks appearance of species in RDA bi-plots of corresponding figures.

686

Fig 1.

754 **Supplementary materials**

755

756 Table A. Number of stingless bee individuals by species captured in the state of 757 Rondônia, Brazil, by the substrate or method of capture.

Trigona sp. n.					7	11	65		$\overline{4}$		3	91	0.95
Trigonisca bidentata Albuquerque & Camargo, 2007							$\overline{2}$		1		23	26	0.27
Trigonisca fraissei (Friese, 1901)									1		20	21	0.22
Trigonisca graeffei (Friese, 1901)					3						13	16	0.17
Trigonisca hirticornis Albuquerque & Camargo, 2007					1		9				15	25	0.26
Trigonisca variegatifrons Albuquerque & Camargo, 2007											1	1	0.01
Trigonisca sp. 1					\overline{c}		5				$\overline{4}$	11	0.12
Trigonisca sp. 2							4		$\overline{2}$		62	69	0.72
Trigonisca sp. 3							1		2		30	33	0.35
Trigonisca sp. 4											$\overline{4}$	4	0.04
Trigonisca sp. 6							2				10	12	0.13
Trigonisca sp. 7											2	3	0.03
No. of species by substrate	13	27	15	14	78	61	79	31	47	18	79	9555	100

759 w=water, m=mud, f=flesh, fec=feces, fl=flower, eu=euglossine bait, h=honey bait, n=nest, o=other,
760 r=resin, s=skin. Species listed in bold are new records for the state of Rondônia. r=resin, s=skin. Species listed in bold are new records for the state of Rondônia.

761

762

763 Table B. List of stingless bee species according to the meso-region in which they were
764 found in Rondônia state, Brazil. An "X" indicates at least one individual found in the

764 found in Rondônia state, Brazil. An "X" indicates at least one individual found in the meso-region.

767 CN-Campo Novo; CM-Costa Marques; EX-Extrema; MC-Machadinho-Cujubim; SM-São Miguel do

Terres, CA-Campo Novo; CM-Costa Marques; EX-Extrema; MC-Machadinho-Cujubim; SM-São Miguel do

768 Guaporé; CÃ-Estação Ecológica Cuniã; PEGM-Parque Estadual Guajará-Mirim; RESEX RCAU-Reserva

769 Extrativista do Rio Cautári Extrativista do Rio Cautário; RESEX ROP-Reserva Extrativista do Rio Ouro Preto; CH-Chupinguaia; LD-

770 Linha D; OP-Ouro Preto do Oeste; PM-Pimenteiras; PVH-BR-364-Porto Velho BR-364; RM-Rolim de
771 Moura. Light shading=newer settlement; dark shading=area of preservation; no shading=older settlemer Moura. Light shading=newer settlement; dark shading=area of preservation; no shading=older settlements.

772

774 Table C. Count of times each stingless bee species was found in each location according

775 to the level of deforestation within 0.5 km of sample location in Rondônia state, Brazil.

776 Number of locations sampled at corresponding deforestation level in parentheses. Dark shading=species
777 found only in locations of lower deforestation levels (<80%); light shading=species found in all locations
778 r found only in locations of lower deforestation levels (<80%); light shading=species found in all locations, regardless of deforestation level; no shading=other.

779 780 Table D. List of stingless bee species and number of locations in which
781 they were found, grouped by whether found in open or closed canopy

they were found, grouped by whether found in open or closed canopy

782 vegetation sub-locations in Rondônia state, Brazil.

783 Light shading=found only in open canopy vegetation; dark shading=found only in

784 closed canopy vegetation; no shading=found in both.

786

787 Table E. Stingless bee species found in nests, grouped by sample sub-location type (open 788 or closed canopy) in Rondônia state, Brazil. or closed canopy) in Rondônia state, Brazil.

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790 Light shading=found only in open canopy; dark shading=found only in closed canopy; no shading=found in

791 both open and closed canopy. 790
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Potential taxonomic issues regarding list of species

- There are a few taxonomic issues in the species list we generated. According to Camargo
- and Pedro (2007), the status of *Oxytrigona flaveola* (Friese, 1900) remains unclear; it
- could be comprised of as many as three species, considering that the type series is
- composed of individuals from Brazil (Espírito Santo), Colombia and Guatemala. The
- situation of *Trigona fulviventris* Guerin, 1835 is also complicated; Camargo and Pedro
- (2007) assume that this species extends from Mexico to western Ecuador, and that there
- are many undescribed species in the group. Moreover, they note that *T. guianae*
- Cockerell, 1910 and *T. braueri* Friese, 1900 have been treated in the literature as *T.*
- *fulviventris*. So, it is possible that the bees identified as *T. fulviventris* in the present study
- are really *T. guianae*, since this occurs in Rondônia and *T. braueri* does not.